some survive in the desert even during years of drought, without any artificial irrigation with only 80 mm of rainfall, and in the subtropical zone with 200-300 mm of rain.

These results and those obtained earlier (Gindel 1968b, 1969) indicate that further study is needed to clarify to what extent augmented stomatal density and the presence of stomata on the upper epidermis in woody xerophytes grown in the driest climate and poorest soil serve to increase the ability of the tree to absorb atmospheric moisture, when moisture tension in the soil and within the plant reach their highest values.

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LITERATURE CITED

Fahn, A. 1964. Some anatomical adaptations of desert plants. Phytomorphology 14: 93-102.

Gindel, I. 1952. Some anatomical features of the indigenous woody vegetation in Israel. Res. Coun. Israel 2: 1-16.

. 1957. Acclimatization of exotic woody plants in Israel: The theory of phytoplasticity. Materiae Vegetabiles 11: 81-101.

. 1961. The water balance in xerophytes. Intern. Union For. Res. Org. 13 Congress 21: 2-4.

-. 1964a. Seasonal flucuations in soil moisture un-

der the canopy of xerophytes and in open areas. Commonwealth Forest. Rev. 43: 219-234.

-, 1964b. Transpiration of Aleppo pine as a function of environment. Ecology 45: 868-873.

-. 1966. Attraction of atmospheric moisture by woody xerophytes in arid climates. Commonwealth Forest. Rev. **45**: 297–321.

—. 1967a. The correlation between cambium activity

and transpiration rate in Aleppo pine. 14th I U F R O Congress 4: 188-207.

The relationship between transpiration . 1967b. and six ecological factors. Oecol. Plant. 8: 227-239. Afforestation of the desert. –. 1967*c*.

Jerusalem 19: 13-24.

-. 1968a. Some ecophysiological properties of three tree xerophytes grown in the desert. Oecol. Plant. 3: 49-67.

-. 1968b. Dynamic modifications in alfalfa leaves growing in subtropical conditions. Physiol. Plant. 21: 1287-1295.

. 1969. Stomata constellation in the leaves of cotton, alfalfa, maize and wheat plants as a function of soil moisture and environment. Physiol. Plant. (in

Kohl, F. G. 1886. Transpiration der Pflanzen und ihre Auswirkung auf die Ausbildung pflanzlicher Gewebe. H. Bruhn, Braunsweig. 34 p.

Sorauer, H. 1873. Einfluss der Wasserzufuhr auf die Ausbildung der Gerstenpflanzen. Bot. Z. 31: 145-159. Stocker, O. 1960. Physiological and morphological

changes in plants due to water deficiency. Arid Zone Research, UNESCO, 15: 63-104.

Zelitch, I. 1961. Biochemical control of stomatal opening in leaves. Proc. Nat. Acad. Sci. Wash. 47: 1423-83.

EXPERIMENTAL ZOOGEOGRAPHY OF ISLANDS: DEFAUNATION AND MONITORING TECHNIQUES

EDWARD O. WILSON AND DANIEL S. SIMBERLOFF¹

The Biological Laboratories, Harvard University, Cambridge, Massachusetts 02138 (Accepted for publication December 16, 1968)

Abstract. In order to facilitate experiments on colonization, a technique was developed that permits the removal of the faunas of very small islands. The islands are covered by a tent and fumigated with methyl bromide at concentrations that are lethal to arthropods but not to the plants.

Seven islands in Florida Bay, of varying distance and direction from immigrant sources, were censused exhaustively. The small size (diameter 11-18 m) and ecological simplicity of these islands, which consist solely of red mangrove trees (Rhizophora mangle) with no supratidal ground, allowed the location and identification of all resident species. The terrestrial fauna of these islands is composed almost exclusively of arboreal arthropods, with 20-50 species usually present at any given moment. Surveys of these taxa throughout the Florida Keys, with emphasis on the inhabitants of mangrove forests, were made during 1967 in order to estimate the size and composition of the "species pool."

The seven experimental islands were defaunated in late 1966 and early 1967, and the colonists were monitored for 17-20 man-hours every 18 days. Precautions were taken to avoid artificial introductions during the monitoring periods.

Introduction

The first attempt to formulate a quantitative theory which would unite an ever-increasing mass of insular biogeographic data was made by Mac-

¹ Present address: Department of Biological Science, Florida State University, Tallahassee, Florida 32306

Arthur and Wilson (1963, 1967), who postulated an equilibrium number of species on an island determined by the intersection of immigration and extinction curves drawn as a function of the number of species already present. In addition to speculation on the forms of these curves and effects of varying both island area and distance from the source, they derived the equation

$$t_{0.90} = 1.15 \frac{\text{mean } S}{X}$$
 (1)

where S_i = the hypothetical equilibrium number of species on an island

mean \check{S} = the mean of \check{S}_i for a number of very similar islands i

X = the extinction rate (or turnover rate, since an equilibrium requires that extinction = immigration) in number of species per time at equilibrium

 $t_{0.90}$ = time required for number of species present to increase from 0 to 90% of \check{S} .

Later the same authors elaborated upon the general shapes of immigration and extinction curves, discussed colonization rates and curves, and predicted the distribution of survival times of species that succeed in colonizing an island (MacArthur and Wilson 1967).

Data to test these hypotheses are scarce. Most biogeographic information, when combined with information furnished by the scant fossil record, can at best lead to descriptions of broad patterns of distribution and zoogeographic regions and to hypotheses about the essentially long-term processes responsible for them (e.g., Darlington 1957). Such theories, though occasionally suggestive for specific groups, fail to explain the

existing taxonomic distributions and the underlying colonization mechanisms for most islands. They neglect the short-term, even daily events that largely determine the parameters of colonization on many islands. Such events become decisively more important than evolution as distance of island from source area decreases. In particular, the classical zoogeographic methods do not provide a test for the existence of an equilibrium number of species or for the accuracy of equation (1).

The literature on biotic dispersal, both anecdotal and systematic, provides information on the relative importance of active and passive transport and a wealth of data on the specific agents of passive transport (e.g., Wolfenbarger 1946). Certain cases of overseas dispersal have even been traced to specific meteorological events (French 1964). For no island, however, is the time course of colonization by even a large fraction of the inhabitants known. The large number of records of long distance dispersal implies that immigration rates to islands are probably high, especially for organisms capable of passive transport by wind. But the shape and determinants of the immigration curve cannot be deduced. At best a lower limit might be given.

In the recent studies of Surtsey, a new volcanic island near Iceland, there has been a deliberate attempt to document the colonization process from

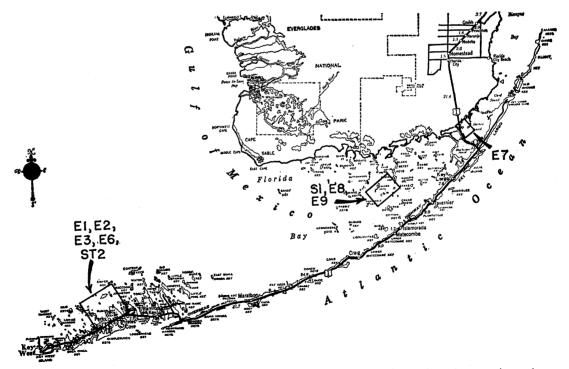


Fig. 1. The southern tip of Florida and the Florida Keys. The rectangles enclose the experimental areas shown in detail in Figures 3-5.

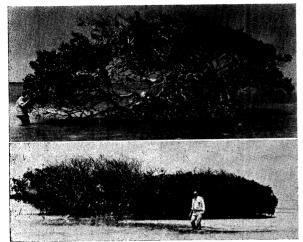


Fig. 2. Upper: Island E1, the second smallest island in the experimental series. Lower: Island E9, the largest island in the experimental series; note also the presence of supratidal mud.

the birth of the island (Fridriksson 1967, Hermannsson 1967, Lindroth et al. 1967). But the infrequent natural occurrence of such events tends to render quantitative biogeographic theory the product of abstract speculation and of enlargement and sophistication of untestable hypotheses. What is needed, clearly, is a method either of producing new islands similar to natural ones or of sterilizing preexisting islands.

THE EXPERIMENTAL ISLANDS

Along the Overseas Highway of southern Florida (U. S. Route 1) there exist a vast array of small, approximately circular islands situated in shallow bay water (Fig. 1). Together, they constitute a potential natural "laboratory" for experimental biogeography. The islands with diameter 10–20 m (Fig. 2) are 5–10 m tall and usually

consist entirely of one to several red mangrove trees (*Rhizophora mangle*), with rarely a small black mangrove bush (*Avicennia nitida*). Occasionally, in areas of weak tide, the trees are surrounded by small areas of supratidal mud and sand. Such land is only intermittently supratidal, since it is completely flooded during prolonged winds of 10 knots or more. Individual islands may differ in several minor respects such as the number of large trunks and amount of dead bark they contain, but on the whole they are remarkably similar to each other in physical appearance.

A fruitful defaunation experiment with these islands requires the following conditions:

- i) that there be enough of the islands for replication and variation in distance to the nearest source area;
- ii) that the animal diversity be sufficient to allow statistical treatment, and the organisms be physically large enough to insure recording inconspicuous forms as well as conspicuous ones;
- iii) that the extremely small size of the islands compensates for their relative nearness to source areas and therefore produces a distance effect.

During an exploratory trip to the Florida Keys in 1966, we found that conveniently accessible small islands were numerous at most distances up to 200 m from the nearest large islands. Beyond 200 m they were rare, although somewhat larger islands (diameter 50 m and greater) were plentiful. A very few small islands suitable for our purposes were located 0.2–1.4 km from the nearest source area.

A set of such islands were chosen for experimentation and their faunas carefully surveyed. Since the breeding fauna of islands this size consists almost entirely of species of insects and spiders, we made a reference collection of land ar-

TABLE 1. Parameters of experimental islands

Series 1	Island name E1 E3 ST2 E2	Diameter (m) 11 12 11 12 11	Distance from nearest source (m) 533 172 154 2	Initial no. arthropod species 25 31 29 43	Location off 'Squirrel Key in Rattlesnake Lumps off Saddlebunch Key off Snipe Keys
	E6s	15ª	73ª	30a	in Johnston Key Mangroves
	E7	25	15	29	near Manatee Creek
	E8	18	1,188 ^b	29	between Calusa Keys and Bob Allen Keys
Series 2	E9	18	379	41	between Calusa Keys and Bob Allen Keys
	E10a	11a	37ª	20ª	near Bottle Key

aControl island b"Stepping Stone" (SI) 572 m away.

thropods of the Florida Keys with the assistance of Robert Silberglied. This collection enabled us to identify the arthropods on the experimental islands, and it provided a measure of the size and composition of the species pool of the presumed source area, the entire Florida Keys.

Animal diversity on the small mangrove islands proved to be adequate for statistical purposes. About 75 insect species (of an estimated 500 that inhabit mangrove swamps and an estimated total of 4,000 that inhabit all the Keys) commonly live on these small islands. There are also 15 species of spiders (of a total Keys complement of perhaps 125 species) and a few scorpions, pseudoscorpions, centipedes, millipedes, and arboreal isopods. At any given moment 20–40 species of insects and 2–10 species of spiders exist on each of the islands.

Many birds visit the islands either to roost or to forage, but very few nest there. The only birds that bred on our islands during the course of the experiment (1966–68) were one or two pairs each of the Green Heron, White-crowned Pigeon, and Gray Kingbird. Snakes (mostly water snakes of the genus *Natrix*) occasionally swim to small mangrove islands, and raccoons visit islands located on shallow mud banks. These vertebrates were not included in the censuses.

On the basis of their distance and direction from the nearest source area, their size, and their accessibility, the islands listed in Table 1 were chosen for defaunation. Two islands (E6 and E10), one each in the vicinity of the two groups of experimental islands, were selected to serve as control islands; and they were censused at the same time as the experimental islands. In order to ascertain whether any long-range change had occurred in the general fauna of small mangrove islands in Florida Bay during the course of the

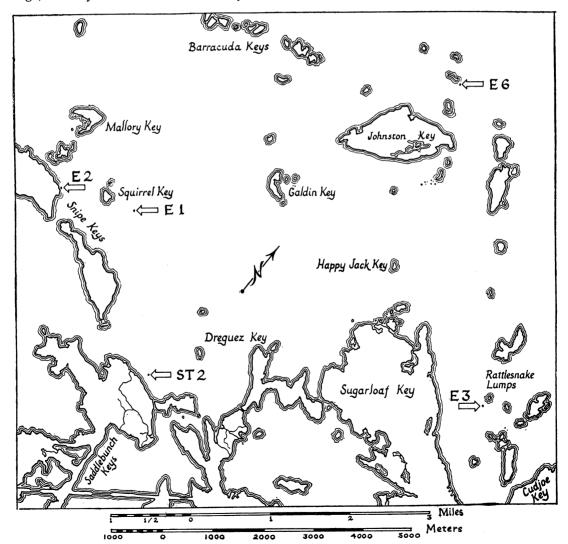


Fig. 3. Map showing the location of the experimental islands of series 1.

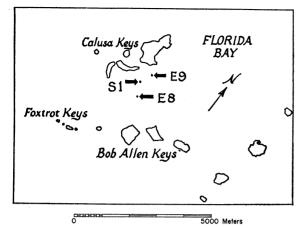


Fig. 4. Map showing the location of the experimental islands of series 2.

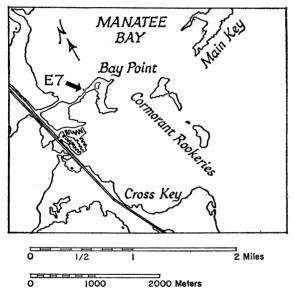


Fig. 5. Map showing the location of the original test island E7.

experiment, a second census of the control islands was made at the close of the experiment.

The two series of experimental islands form two widely separated groups. Series 1 lies within the Great White Heron National Wildlife Refuge, in a region north of Sugarloaf Key (Fig. 3). Series 2 is in the Everglades National Park near the Calusa Keys north of Islamorada (Fig. 4). E7, which was chosen primarily to test a method of defaunation, is located in Barnes Sound south of the Glades Canal (Fig. 5).

The islands were also selected to show as much variation as possible in the degree of isolation from the immigrant sources. The coefficient of correlation for island series 1 between "distance from nearest source" and "number of arthropod species" is —0.83, indicating the operation in the

original faunas of the distance effect as defined by MacArthur and Wilson (1967). Qualitative differences in faunas of increasingly distant islands were even more striking than quantitative ones. For many groups—especially ants and spiders—the distance effect is so regular that one can guess not only the species number but also the approximate species composition. For example, centipedes and pseudoscorpions were found almost exclusively on islands of less than 200 m distance from the nearest large island.

Islands of this minute size can be remarkably long lived. All Rhizophora mangle trees, and particularly those on low, small islands, have numerous aerial roots: both brace roots growing out from the main stem and prop roots growing down from branches. The latter can emerge from as high as 4 m above the water (LaRue and Muzik 1954) and in time can become as thick as the main stem. Rhizophora trees reproduce viviparously, their seeds growing on the tree until they become large, long, pointed seedlings. When they drop, frequently from the upper canopy, some evidently plant themselves where they fall (Davis 1940). Mangrove swamps are often quite thick, and on small islands there may be several trees. Stems, thickened prop roots, and brace roots in-

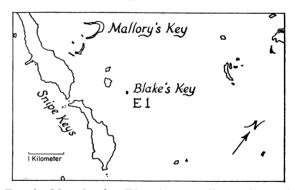


Fig. 6. Map showing E1 and surrounding region as it appeared in 1851-57.

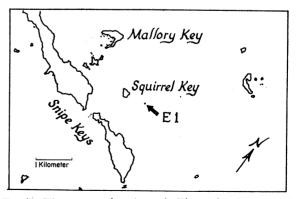


Fig. 7. The same region shown in Figure 6 as it appeared in 1964.

termingle and reproduce themselves so that it is impossible to distinguish how many trees are present, which are which, which is the initial one (even if it is still present), and which are the main stems.

Maps of the Florida Keys, especially of the Sugarloaf Key-Key West area, have been remarkably detailed and accurate with respect to small mangrove islands for at least a century. These tell us that E7 and all the islands in series 1 have existed at least 25 years. The other experimental islands were not mapped in sufficient detail to determine longevity. The size and shape of individual islands may change, even drastically, but in general the islands survive the climatic catastrophes to which they are repeatedly subjected. For example, Figure 6 shows E1 (Blake's Key) and surrounding areas as it appeared in 1851-57, while Figure 7 shows the same island region on the Coast and Geodetic Survey map of 1964. It therefore seems reasonable to assume that the faunas found initially on the experimental islands were the product of many years' colonization. Our experiments have subsequently shown that the minimum ages of the islands greatly exceed the time required to reach species equilibrium from a start of zero species (Simberloff and Wilson 1969).

DEFAUNATION

Our first attempt to defaunate islands made use of an insecticide spray of short-lived residual effect. On July 9-10, 1966, two experimental islands (E1 and E2) were sprayed until dripping with 60 g parathion, 240 g diazinon and 180 g Pylac sticking agent per 100 l fresh water. When the islands were examined 1 day later, all the surface and bark fauna and most borers were dead. However, the following live animals were found in thin hollow twigs. E1: one adult wasp, Scleroderma macrogaster (Bethylidae); workers and larvae of Crematogaster ashmeadi (Formicidae). E2: two larvae of ?Styloleptus biustus (Cerambycidae); two colonies of Paracryptocerus varians (Formicidae); and one colony of Tapinoma littorale (Formicidae). A more thorough examination 9 days after spraying produced the following live and apparently healthy insects. E1: two larvae of ?Styloleptus biustus (Cerambycidae). E2: one larvae of ?Styloleptus biustus (Cerambycidae); one lepidopteran larva; and two colonies of Xenomyrmex floridanus (Formicidae). The large populations of the ant colonies and advanced development of the beetle larvae precluded these insects having immigrated during the 9 days following spraying.

The results on E1 and E2 showed that a spray

cannot be expected to penetrate all the hollow twigs of a *Rhizophora* island. Some twig dwellers will probably survive, particularly if they inhabit narrow twigs (*Tapinoma* and *Xenomyrmex*) or else pack their tunnels tightly with powdery excreta (*Styloleptus*). Had we attempted to kill these survivors by using a spray with a long-lived or more powerful residue, we would have postponed the beginning of the colonization curve, since immigrants may be killed by the residue. Because we had to be certain that all colonists, or at least all but those belonging to a small number of known taxa, would be destroyed immediately, we abandoned the spray technique.

We turned next to the more difficult alternative technique of fumigation. The method has the obvious advantages that the residual effect of a gas is negligible, and even the inhabitants of hollow twigs are contacted. This is the standard method used by professional exterminators on termites and other wood-boring insect pests.

The fumigation of living plants is a relatively uncharted area. The biochemical effects of fumigants on plants are poorly known, damage usually being described in terms of the physical appearance of the plant shortly after fumigation (Page and Lubatti 1963). To our knowledge, the only prior fumigation of red mangrove trees was by B. P. Stewart (pers. comm.), who used methyl bromide on small plants. With 32 kg/1000 m³ for 2 hr at 32°C and 40 kg/1000 m3 for 2 hr at 27°C he found that all insects apparently were killed, although some did not die until 48 hr after Mortality of cockroach eggs was fumigation. unrecorded in Stewart's study. Damage to the plant was limited primarily to young leaves and twigs, a result that is in accord with the general finding of Page and Lubatti (1963) that the growing stages of both plants and insects are the most susceptible to fumigation.

In our preliminary trials we were limited by the fact that fumigants highly soluble in water, such as hydrogen cyanide, could not be used. These substances would be expected either to leave the fumigation tent through the water or to concentrate to an unknown extent in the water Either effect would greatly under the island. complicate the maintenance and measurement of gas concentration. Consequently, the following four relatively insoluble fumigants were tried: Pestmaster Soil Fumigant-1 (methyl bromide 98% wt, chloropicrin 2% wt); Acritet 34-66 (acrylonitrile 34% vol, carbon tetrachloride 66% vol); Dow Ethylene Oxide (ethylene oxide 100%); and Vikane (sulfuryl fluoride 95% wt, inert ingredients 5% wt). Field tests were performed on small Rhizophora trees in Matheson Hammock Park, Coral Gables, by National Exterminators of Miami, Florida, under the direction of Steve Tendrich in consultation with the authors. Additional tests on the mortality of mangrove colonists were performed at National Exterminators' laboratory in Miami.

With the temperature outside the fumigation tent remaining at 24°-29°C, we determined which duration-concentration intensities caused no more than bearable damage to the tree yet were lethal to all the mangrove arthropods. Acritet, Vikane, and ethylene oxide caused extensive, apparently irreversible damage to red mangrove and were However, methyl bromide at 22 kg/ 1000 m³ for 2 hr caused browning of only about 10% of the leaves 1 week after treatment, with no visible later effects. This duration and concentration killed all the mangrove inhabitants (including such resistant forms as roach eggs and lepidopteran pupae) with one possible exception. When data from all tests at this duration-concentration are lumped together, only 95% of cerambycids and 80% of weevils were dead when dug out of their burrows 1-10 days after fumigation. The remainder all died within 2 days of removal. It is tempting to ascribe the death of these few larvae to a delayed effect of the fumigant. However, a control experiment consisting of the removal of five unfumigated cerambycid larvae (Styloleptus biustus) from their respective burrows caused death within 5 days without any additional treatment. Monro and Delisle (1943) report that many insects are active soon after exposure to methyl bromide but succumb in time (some lepidopterous larvae surviving for 2 months), and that this delayed effect is especially common in fumigations at the low concentrations used in our own field tests. This peculiarity is confirmed by Chisholm (1952) for Japanese beetle larvae and other insects. Of at least equal importance, the results from our experimental islands (Simberloff and Wilson 1969) strongly suggest that the cerambycids were eliminated by the fumigation treatment. We therefore conclude that even the deep-boring beetle larvae were very probably killed by the treatment.

The overt physical damage to the trees consisted of a browning of leaves and shoots and occasional heavy oozing of sap. At 61 kg/1000 m³ the leaf browning was manifest upon tent removal, but at 35 kg/1000 m³ and less there was little or no immediate browning. The effect became evident at about 24 hr and worsened in terms of fraction of leaves browned through 72 hr, by which time up to 90% of the leaves were brown and after which no further browning occurred. Browning did not occur randomly throughout the

canopy. Instead, whole sections of the tree became completely brown while others remained almost unscathed. The upper canopy was always more severely browned, but there seemed to be no other consistent damage pattern.

Abscission of the dead leaves began within a week and, often aided by wind, was usually complete within 8 weeks. At this time damaged trees looked superficially dead except for remaining green sections. Closer examination showed that the damaged trees could be conveniently divided into three classes, as follows:

For those trees where browning of leaves and shoots appeared complete by 72 hr the damage was irreversible. Within a few months bark began to peel and twigs with green wood could not be found. A year later there was no evidence of life. Island E8 is in this class. In the second category, which includes only E7, a large majority of the leaves and shoots were browned and died, while a small section remained green. In the third category, the trees were less severely damaged (60% or fewer browned leaves). Green sections remained unchanged, while leaf abscission in most of the damaged sections was accompanied by the sprouting of new shoots and leaves from almost all branches. Within 2 months the trees appeared normal except for occasional leafless patches in the extreme upper canopy. Island E1, which 2 days after fumigation appeared to have about 50% brown leaves, recovered after 1 month to the extent that it did not appear to have incurred more damage than similar trees do during heavy storms, and 2 months after defaunation seemed quite normal except for a small bare patch in the upper canopy. At no time did its bark crack or peel.

In most damaged trees of the first two classes—where either the whole tree or all but a small section was killed—there was copious oozing of sap for up to a week after fumigation. Such oozing was never observed in untreated trees, nor in those fumigated trees which showed good leaf replacement.

The action of methyl bromide on living plants has been studied rather extensively by several authors. Mainwaring (1961) examined the physiological responses of several plants to methyl bromide, and his numerous observations imply that its major effect is auxin inhibition. The damage we observed seemed too drastic and certainly too immediate to be attributed solely to auxin inhibition, although the presence of such a biochemical action cannot be ruled out.

Whatever the effect of methyl bromide on Rhizophora mangle, it apparently had the expected Q_{10} of 2-3. Our initial tests in Matheson Hammock were conducted on small plants well shaded

by large trees, and the temperature inside the fumigation tent remained within 3° of the air temperature (24°-29°C). The experimental islands could not be shaded, and even on overcast days the temperature inside the dark tent was at least 32°C. On sunny days the inner temperature was so high that the presence of the covering alone caused browning of leaves. It seems probable that the generally more severe damage to the upper canopy throughout this experiment was caused by higher temperature, especially since methyl bromide gas has a density (3.974 g/liter at 20°C) considerably greater than that of air.

Air humidity is much less important than temperature in its effect on fumigation damage to plants (Page and Lubatti 1963); our experiments were almost all conducted within a narrow humidity range (65%–90%) and the effects of varying this parameter were not studied.

The first two experimental islands (E7 and E1) were fumigated during daytime and damaged. In order to reduce heat damage we therefore fumigated all other islands at night. The concentration was increased from 22 kg/1000 m³ to 28–30 kg/1000 m³ with no observable additional damage.

The chemical means for defaunation now having been chosen and tested, we were left with an overwhelming physical problem of how to fumigate an entire island. Assuming that we could somehow get the fumigation tent (a 25-m by 25-m piece of plastic-impregnated canvas with addi-

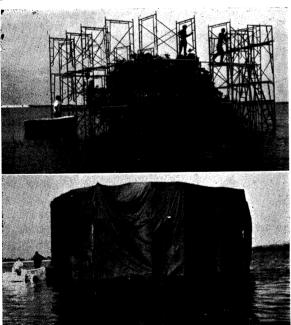


Fig. 8. Upper: The scaffolding constructed around E7, complete except for the top walkway. Lower: the fumigation tent over E7.

Table 2. Arthropod species found on E7 just prior to and following fumigation

NSECTS			
Embioptera:	gen. sp.a		
Orthoptera:	$Latiblattella \ { m n. \ sp.}$		
	$Cycloptilum \mathrm{sp.}$		
	$Tafalisca\ lurida^{a}$		
Coleoptera:	$Pseudoacalles m ~sp.^b$		
1	Leptostylus sp.		
	Styloleptus' biustus ^b		
	Tricorynus sp.		
Psocoptera:	Archipsocus panama		
Hemiptera:	Pseudococcus sp.		
Lepidoptera:	Nemapogon sp.a		
Hymenoptera:	Casinaria texana		
F	Camponotus floridanusa		
	Camponotus planatusa		
	Paracryptocerus variansa		
	Pseudomyrmex elongatus ^a		
	Xenomyrmex floridarus		
D A CHINTIDO			
RACHNIDS	And a face must seed		
Araneae:	Ariadna arthuri		
	Eustala sp.		
	Leucauge venusta ^a		
Acarina:	Galumna sp.		
THER			
Chilopoda:	Orphnaeus brasilianusa		
Isopoda:	Rhyscotus sp.a		

adiscovered dead after fumigation bdiscovered mostly dead after fumigation (see text)

tional side pieces for larger islands) onto an island, the tree could not support the weight without major limb breakage, since even our special half-weight tent weighed 275 kg. It was evident that some sort of superstructure was needed, one which would surround and rise over the island to remove most or all of the tent weight from the tree. Constructing a permanent framework and lowering it over the islands by helicopter was too expensive.

An alternative method employing a temporary full framework constructed on the site was used on October 10, 1966 to defaunate E7. The procedure was as follows (see also Fig. 8): First, planks to prevent the framework from sinking into the mud were placed underwater in a square around the islands and held manually until the first tier of scaffolding (which was sufficiently heavy to keep them from floating) was placed on top. Then the remaining tiers were added, one side at a time, until the island was surrounded by a cube of scaffolding. Fig. 8 (upper) shows the framework complete except for the upper tier of scaffolding in the right rear corner and the square walkway of planks around the very top of the structure.

The tent was then raised by block and tackle to the top of the framework and lowered over the island. Methyl bromide was introduced from 45 kg cylinders through the tent wall at opposite sides of the island, 2.5 m above the water, and into a metal tub mounted with an electric fan to disperse the gas. Chloropicrin (tear gas), added to the odorless methyl bromide as a safety measure, was allowed to drip into burlap sheets at the bottom of the tubs. Concentration of the methyl bromide was measured frequently by a Gow-Mac thermal conductivity meter with a self-contained pump attached to a testing station 2.5 m above the water and well away from the two shooting stations.

The methyl bromide concentration was gradually built up from 0 to 22 kg/1000 m³ over 30 min, then kept at 22–25 kg/1000 m³ for 2 hr. Outside air temperature during this time remained steady at about 27°C. At the end of the fumigation the gas was released through seam openings for 45 min, and the tent was then removed.

Immediately following the dissipation of the gas, E7 was explored for a total of 6 man-hours. The results generally confirmed the evidence from the Matheson Hammock tests. Dead individuals representing 11 of the 23 species observed before fumigation were found, frequently in large numbers and in all life stages. The details are given in Table 2. The only live animals encountered were larvae of two deep-boring beetles. formed only a small fraction of the original population, as shown by the following recovery data: of 59 larvae of Styloleptus biustus (Cerambycidae), 2 were alive and 57 dead; of 10 larvae of Pseudoacalles sp. (Curculionidae), 2 were alive and 8 dead. One of the two live cerambycids was extremely sluggish and died an hour after removal. The other cerambycid and the two weevils seemed healthy, though one of the weevils became torpid and died 9 hr after removal.

The tree was severely damaged as described previously. Within 2 months 85% of the foliage was dead, although the single living section has remained green until at least May 1968. Thus the procedure worked essentially as expected, and this aspect of the fumigation was left unchanged in subsequent defaunations, except that the work was henceforth done at night in all but one instance (E1). The cubical scaffolding framework had disadvantages, chief among them being the large crew necessary to erect it (at least five men) and the number of pieces and weight of the equipment. This necessitated numerous boat trips between E7 and our staging area near Manatee Creek, since the shallow water along the route precluded the use of large craft. This problem was greatly aggravated on the other islands, which are located considerably further from a convenient staging area. We therefore set out to devise a tower to be located in the center of the island and rising high enough so that most of the tent weight rested on it. This device proved to be more vul-

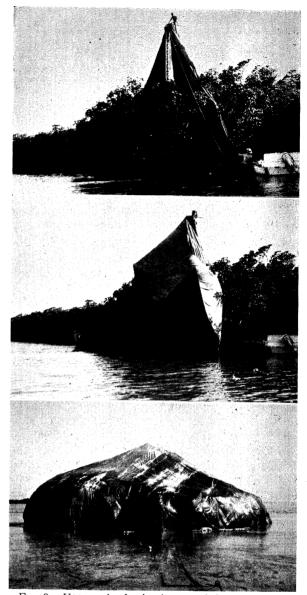


Fig. 9. *Upper*: the fumigation tent being opened over E2 from a tower support. *Middle*: E2 partially covered. *Lower*: E9 covered by the fumigation tent on a tower support.

nerable to high winds, but it has the great advantage of requiring less equipment and a smaller crew to assemble it. National Exterminators contracted a steeplejack to perfect the tower method. After a month of experimenting on islands near Key Largo, the following procedure was adopted and used for the remaining islands.

A collapsible 10-m triangular steel tower was erected in the center of the island, either propped on a large root or forced deep into the ground. The steeplejack then climbed to the top and threw out three guy wires which were fastened down

TABLE 3. Extent of damage to foliage and subsequent recovery on experimental islands

Island	Fumigation date	Time	Extent of damage (burnt leaves)	Damage location	Recovery of foliage
E1	3/13/67	day	50%	general	complete in 2 months, except for extreme upper canopy
E2 E3	$\frac{3/7/67}{3/17/67}$ $\frac{10/10/66}{10}$	night night day	10% 5% 85%	upper canopy upper canopy all except one	complete in 2 months complete in 2 months none
E8 E9	4/17/67 4/7/67	night night	100%	section complete upper canopy and one other	none little in upper canopy complete elsewhere in
ST2	4/20/67	night	5%	section scattered	2 months complete in 2 months

with mudscrews or stakes, depending on the substrate. Using a block and tackle, one man on the tower and one on the island lifted the rolled tent until it formed a triangle over the island. The tent was then carefully unrolled down the guide wires until the island was covered (Fig. 9). It was impossible to unfurl the tent from the tower in winds greater than 8 knots. Sandbags and stakes were used to keep the tent edges submerged.

Equipment used in the tower method was identical to that described already for the scaffolding method, and was similarly situated. The procedure was identical except that gas was administered from 0.5-kg cans and concentration and duration were increased to 28-33 kg/1000 m³ for 2.5 hours on all islands but E1. After fumigation the methyl bromide was released through tent seams as with scaffolding, but in the tower system tent removal was more difficult. At low wind speeds the tent could be laboriously refurled and gradually lowered to the boat. We fortuitously discovered that high winds, instead of causing extensive damage as we expected, could actually aid tent removal. The windward side of the tent was lifted by poles to about 3 m above water, whereupon sufficient wind entered the tent to lift it up over the tree and to drop it, partly extended, into the water on the leeward side of the island.

E1 was fumigated in the daytime. For all remaining islands the tent was unrolled no earlier than 60 min before sundown, the fumigation was conducted between 8:00 and 12:00 pm, the gas was removed by 1:00 am, and the tent was either off the island or refurled within 90 min of dawn. Effects on the mangrove trees of the fumigated islands are noted in Table 3.

All islands were examined immediately after tent removal. No living animals at all were found on E1, E2, E8, and E9. A single live curculionid larva was found on E3. On ST2, 1 living and at least 100 dead polyxenid millipedes (Lopho-

proctinus ?bartschi) were found. Although we have no further direct evidence and know of no work on methyl bromide fumigation of millipedes, we suspect that all millipedes not destroyed immediately were killed by the delayed effect of methyl bromide discussed earlier.

In sum, we feel that this fumigation technique killed all mangrove arthropods with the single remotely possible exception of deep-boring beetle larvae.

Monitoring Technique

Precautions were taken at all monitorings² to prevent contamination of the experimental islands. Boats were never brought in contact with the islands and only seldom tied to them; usually they were anchored at least 12 m away. Before wading to an island, experimenters examined clothing, equipment, and persons for animals, which were destroyed. To lessen the chance of phoresy, investigators simply immersed themselves briefly or sprayed themselves with "Off" insect repellent. All equipment used on the island was sprayed weekly with a short-lived insecticide.

The absence of supratidal land, except around E9, greatly simplifies the environment of the experimental islands. The arboreal substrate may be arbitrarily divided into hollow twigs, living branches and twigs, dead bark and tree holes, leaves and flowers, green shoots, and fruits. Because of the extremely small size of the experimental islands it was possible to study most of these microhabitats all but exhaustively for the pre-defaunation censuses. A typical example is the survey of E3, where we broke and examined 3,500 of an estimated 5,000 small hollow twigs, collected

² The term "monitoring" is used to designate one or more of the censuses of species, conducted for a period of 17-20 hr during the course of the experiment. In the present series of articles, monitoring, census, and censusing are used interchangeably to refer to the listing of species.

90% of the fruits, examined virtually all of the leaves, and looked under at least 80% of the pieces of dead bark. *Rhizophora* has numerous limbs and our islands were relatively low, so that we were able, by climbing carefully, to examine the canopy right to the top.

The fumigation procedure itself gave added evidence of the completeness of our original surveys. The fumigant included 2% chloropicrin, which drives out even deep-boring and hollow-twig-dwelling arthropods (cf. Lauck 1965). The animals either fell into the water or remained hanging from branches by spines or silk, tenuously and conspicuously. Those which fell generally remained afloat and, when a tent seam was opened after fumigation, flowed out in masses where they were collected on a screen.

The chloropicrin technique was especially effective on E8, E9, and ST2, and the results indicated greater densities of both insects and spiders than we had suspected. Only on E8, however, were animals found that had not been noted in the original surveys. These consisted of four beetles:

Staphylinidae: gen. sp.

Carabidae: Bembidion sp. nr. contractum

Carabidae: Tachys occulator Oedemeridae: Oxacis sp.

Of these the first three were apparently living in a patch of temporarily supratidal mud, which was submerged again 2 weeks later during a strong wind. The *Oxacis* were encountered only on the fumigation tent, and even their status as "transients" is doubtful.

Defaunated islands were censused for approxi-

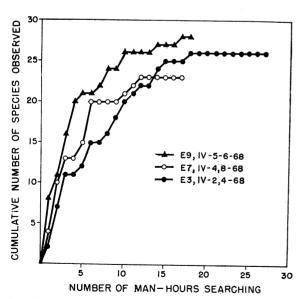


Fig. 10. Cumulative species counts during individual censuses on three of the experimental islands.

mately 2 days every 18 days. At each census period approximately 10% of all hollow twigs were broken, and with this treatment the number of hollow twigs seemed to remain constant throughout the experiment. Only when gentle lifting failed to allow vision was dead bark removed, and then only a part was peeled off. Small amounts of carbon tetrachloride and ammonia, and tapping with a steel probe were used to drive animals out from underneath bark and inside hollow twigs without damage to tree or colonist. though the pre-defaunation searching methods had been drastic, care was taken following defaunation to prevent the censusing from being destructive and from altering the proportions of the respective microhabitats on an island.

Most colonists could be identified in the field because of the small subset of the Keys fauna which invades small mangrove islands and the completeness of our reference collection. Those that could not be recognized with certainty were treated in one of two ways. If the animal in question was obviously part of a sizable population, as with many psocopterans and thrips, a small collection was made that did not exceed 2% of the estimated minimum size of the population. When an unknown arthropod was not part of a large population, we photographed it.

It was quickly discovered that the rate of discovery of new species during each censusing period, although at first very high, declined approximately linearly to near 0 at about 14 man-hours of searching. A very few species are expected to appear from immigration within the monitoring period. Figure 10 depicts the cumulative number of species observed vs. time for three representative monitorings. Different sizes of islands, weather conditions, and especially varying population sizes caused minor differences in the curves from island to island and from time to time; but despite great changes in the numbers of species through the course of the experiment, the curves remained remarkably similar in shape. The censusing period was therefore chosen to last between 17 and 20 man-hours.

Because of the simplicity and very small size of the experimental islands, and because several nocturnal censuses located no new species, we felt that all colonists having crepuscular or nocturnal activity were being recorded during the diurnal monitoring. The anyphaenid spider Aysha sp. and ant Paracryptocerus varians are both strictly nocturnal, yet both were recorded during the day in their retreats. Similarly the crepuscular braconid wasp Macrocentrus sp. and largely nocturnal ants Camponotus floridanus and Tapinoma littorale were repeatedly observed.

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LITERATURE CITED

Chisholm, R. D. 1952. Nature and uses of fumigants, p. 358. In Insects, The Yearbook of Agriculture 1952. Darlington, P. J. 1957. Zoogeography: the geographi-

Darlington, P. J. 1957. Zoogeography: the geographical distribution of animals. Wiley, New York. 675 p.

 Davis, J. H. 1940. The ecology and geologic role of mangroves in Florida. Papers from Tortugas Lab., Carnegie Inst. Wash. 32: 307-412.

French, R. A. 1964 (1965). Long range dispersal of insects in relation to synoptic meteorology. Proc. Int. Congr. Entomol. 12th (London). 6: 418-419.

Fridriksson, S. 1967. Life and its development on the volcanic island, Surtsey. Proc. Surtsey Research Conf., 1967: 7-19.

Hermannsson, S. 1967. Introduction. Surtsey Research Progress Report, III: 1.

LaRue, D. C., and T. T. Muzik. 1954. Growth, regeneration and precocious rooting in *Rhizophora mangle*. Pap. Mich. Acad. Sci. Arts Lett. (I) 39: 9-29.

Lauck, D. R. 1965. Chloropicrin for fast action with a Berlese funnel. Turtox News 43: 115.

Lindroth, C. H., H. Andersen, H. Bodvarsson, and S. H. Richter. 1967. Report on the Surtsey investigation in 1966. Terrestrial invertebrates. Surtsey Research Progress Report, III: 59-67.

MacArthur, R. H., and E. O. Wilson. 1963. An equilibrium theory of insular zoogeography. Evolution

17: 373–387.

MacArthur, R. H., and E. O. Wilson. 1967. The theory of island biogeography. Princeton University Press. 203 p.

Mainwaring, A. P. 1961. Some effects of methyl bromide on aphids and whitefly and their host plants. Doctoral thesis, London Univ., London, England.

Monro, H. A. U., and R. Delisle. 1943. Further applications of methyl bromide as a fumigant. Sci. Agr. 23: 546-556.

Page, A. B. P., and O. F. Lubatti. 1963. Fumigation of insects. Ann. Rev. Entomol. 8: 239-264.

Simberloff, D. S., and E. O. Wilson. 1969. Experimental zoogeography of islands. The colonization of empty islands. Ecology 50: 278-296.

Wolfenbarger, D. O. 1946. Dispersion of small organisms. Amer. Midland Naturalist 35: 1-152.

EXPERIMENTAL ZOOGEOGRAPHY OF ISLANDS: THE COLONIZATION OF EMPTY ISLANDS

DANIEL S. SIMBERLOFF¹ AND EDWARD O. WILSON
The Biological Laboratories, Harvard University, Cambridge, Massachusetts 02138

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Abstract. We report here the first evidence of faunistic equilibrium obtained through controlled, replicated experiments, together with an analysis of the immigration and extinction processes of animal species based on direct observations.

The colonization of six small mangrove islands in Florida Bay by terrestrial arthropods was monitored at frequent intervals for 1 year after removal of the original fauna by methyl bromide fumigation. Both the observed data and climatic considerations imply that seasonality had little effect upon the basic shape of the colonization curves of species present vs. time. By 250 days after defaunation, the faunas of all the islands except the most distant one ("E1") had regained species numbers and composition similar to those of untreated islands even though population densities were still abnormally low. Although early colonists included both weak and strong fliers, the former, particularly psocopterans, were usually the first to produce large populations. Among these same early invaders were the taxa displaying both the highest extinction rates and the greatest variability in species composition on the different islands. Ants, the ecological dominants of mangrove islands, were among the last to colonize, but they did so with the highest degree of predictability.

The colonization curves plus static observations on untreated islands indicate strongly that a dynamic equilibrium number of species exists for any island. We believe the curves are produced by colonization involving little if any interaction, then a gradual decline as interaction becomes important, and finally, a lasting dynamic equilibrium. Equations are given for the early immigration, extinction, and colonization curves.

Dispersal to these islands is predominantly through aerial transport, both active and passive. Extinction of the earliest colonists is probably caused chiefly by such physical factors as drowning or lack of suitable breeding sites and less commonly by competition and predation.

¹ Present address: Department of Biological Science, Florida State University, Tallahassee, Florida 32306